National Parks & Wildlife Service

Camderry Bog SAC (site code 002347)

Conservation objectives supporting document - raised bog habitats

Version 1

November 2015

Contents

1	INTROI	DUCTION	2
	1.1 R	AISED BOGS	2
	1.1.1	Raised Bogs Microtopography	3
	1.1.2	Typical Flora of Irish Raised Bogs	5
	1.1.3	Typical Fauna of Irish Raised Bogs	6
	1.2 H	ABITATS DIRECTIVE RAISED BOG HABITATS IN IRELAND	9
	1.2.1	Restoration of Active Raised Bog in Ireland	10
	1.3 C/	AMDERRY BOG SAC	
	1.3.1	Flora of Camderry Bog	11
	1.3.2	Fauna of Camderry Bog	13
2	CONSE	RVATION OBJECTIVES	14
	2.1 A	REA	14
	2.2 R	ANGE	15
	2.3 S1	RUCTURE AND FUNCTIONS	15
	2.3.1	High bog area	16
	2.3.2	Hydrological regime: water levels	16
	2.3.3	Hydrological regime: flow patterns	17
	2.3.4	Transitional areas between high bog and surrounding mineral soils (include	les
	cutove	rareas)	
	2.3.5	Vegetation quality: central ecotope, active flush, soaks, bog woodland	19
	2.3.6	Vegetation quality: microtopographical features	20
	2.3.7	Vegetation quality: bog moss (Sphagnum) species	20
	2.3.8	Typical ARB species: flora	21
	2.3.9	Typical ARB species: fauna	
	2.3.10	Elements of local distinctiveness	
	2.3.10		
	2.3.10		
	2.3.10 2.3.11	Negative physical indicators	
	2.3.11	Vegetation composition: native negative indicator species	
	2.3.12		
	2.3.13	Vegetation composition: non-native invasive species	
		, , , , , ,	
	2.3.15	Water quality	25
3	REFERE	NCES	26

- Map 1: Extent of potential active raised bog on Camderry Bog.
- Map 2: Distribution of raised bog ecotopes on Camderry Bog.
- Map 3: Digital elevation model and drainage patterns at Camderry Bog.

1 Introduction

This document presents a summary of the background information that has informed the process of setting the Site-Specific Conservation Objective in relation to the priority Annex I habitat 'active raised bog' (habitat code 7110) (hereafter referred to as Active Raised Bog (ARB)), for which Camderry Bog Special Area of Conservation (SAC) has been designated.

Camderry Bog SAC is also designated for two other related Annex I habitats, namely; 'degraded raised bogs still capable of natural regeneration' (habitat code 7120) (hereafter referred to as Degraded Raised Bog (DRB)) and 'depressions on peat substrates of the Rhynchosporion' (habitat code 7150). Based on the close ecological relationship between these three habitats types, it is not necessary to set specific Conservation Objectives for all three habitats individually. It is considered that should favourable conservation condition for ARB be achieved on the site, then, as a consequence, favourable conservation condition for the other two habitats would also be achieved.

1.1 Raised Bogs

Raised bogs are accumulations of deep peat (typically 3-12m) that originated in shallow lake basins or topographic depressions. The name is derived from the elevated surface, or dome, that develops as raised bogs grow upwards through the accumulation of peat; the domed effect is often exaggerated when the margins of a bog are damaged by turf cutting or drainage, and are drying out. Raised bogs are most abundant in the lowlands of central and mid-west Ireland.

Irish raised bogs are classified into two sub-types (Schouten 1984): 1. Western or intermediate raised bogs, and 2. True midland or eastern raised bogs, based on phytosociological and morphological characteristics. In terms of overall morphology, the main difference between these two raised bog types is that while eastern raised bogs tended to stay more confined to the depressions in which they were formed, western raised bogs tended to grow out beyond their original basin, presumably a result of the higher rainfall levels (Cross 1990). In terms of vegetation differences the most obvious difference between the two bog types is the presence of a number of oceanic plant species on western raised bogs which are absent from the true midland raised bogs. The liverwort species *Pleurozia purpurea*¹ and the moss species *Campylopus atrovirens* grow on western raised bogs but not on eastern raised bogs; similarly, *Carex panicea* is generally more common on the high bog surfaces of western raised bogs (Schouten 1984). All of these plant species are widespread in the low-level Atlantic blanket bogs and their presence in western raised bogs is presumed to be due to the higher rainfall levels and greater rain-derived nutrient fluxes.

Exploitation has been extensive and none of the remaining Irish raised bogs are completely intact (Cross 1990). It is estimated that less than 10% of the original raised bog habitat in Ireland is in a near intact state (uncut), with less than 0.5% continuing to support ARB (DAHG 2014). Excavated face banks, whether active or inactive, are a common feature around the margins. Any areas where part of the bog has been removed are termed cutover bog, with the remaining area referred to as high bog or intact bog. In a natural state, raised bogs are circled by a wetland fringe, known as the lagg zone, which is usually characterised by fen

_

¹ Note on species nomenclature: In the case of plant species, only scientific names are used throughout the main text while common English names are included in tables. In the case of faunal species, common English names are used throughout the text (where known) together with scientific names.

communities. In Ireland, most laggs have been lost through drainage and land reclamation (Fossitt 2000).

The surface of a relatively intact raised bog is typically wet, acid, deficient in plant nutrients, and supports specialised plant communities that are low in overall diversity and comprising species adapted to the biologically harsh conditions. The vegetation is open, treeless and bog mosses or *Sphagnum* species dominate the ground layer. Small-scale mosaics of plant communities are characteristic and reflect the complex microtopography of hummocks and hollows on the bog surface (see Section 1.1.1 below). Raised bogs are driest at the margins and wetness generally increases towards the centre of the peat mass where well-developed pool systems are most likely to occur.

Raised bogs may also contain soaks and flushes (wet 'active' or dry 'inactive') due to the increased supply of nutrients over time through concentrated surface flows, or where there are links with regional groundwater or the underlying mineral substratum. Slight mineral enrichment and / or constant through flow of water provide conditions suitable for a range of species that are not typically associated with other areas of raised bog.

When damaged by peat extraction or drainage, the water table in the peat drops and the bog surface becomes relatively dry; pools are rare or absent, cover of bog mosses is greatly reduced and *Calluna vulgaris* increases in abundance. The drop in water table causes the peat to compress under its own weight causing the bog surface to deform. Greater deformation occurs closest to areas where the water table has dropped. This increases the slope of the bog surface causing rain falling on the ground surface to flow off the bog more quickly. The effect is normally greatest around the margins and in a typical situation surface wetness increases towards the centre of the bog. Trees such as *Betula pubescens* and *Pinus sylvestris* frequently invade the drier cut margins, but may also occur in flushed areas.

In Ireland, the Annex I habitat ARB is currently considered to be in unfavourable bad conservation status principally as a result of marginal turf cutting, more recent semi-industrial peat extraction, and associated drainage effects caused by these activities (NPWS 2008; 2013). The lowering of regional groundwater levels are also known to have had an effect on some sites. Fires associated with turf cutting, dumping, or agricultural activities may also adversely affect the condition of the habitat.

1.1.1 Raised Bogs Microtopography

Raised bogs are typically treeless and are characterised by a distinctive vegetation dominated by bog mosses (*Sphagnum*), sedges, and dwarf shrubs, all of which are adapted to waterlogged, acidic and exposed conditions. Bog mosses, which have unique properties, are the principal component of peat, and are largely responsible for the typical surface features of hummocks, hollows, lawns, and pools. The wettest bogs, which have extensive pool systems, have the greatest variety of plant and animal life and support a range of specialist species.

The following terms that describe microtopography are generally accepted in the study of mire ecology (Gore 1983). A schematic diagram showing the typical microtopographical divisions is presented in Figure 1.

Pools

Depressions in the bog surface where the water table remains above the surface level all year around or below surface level for only a very short period of time. They are characterised by the presence of aquatic plant species such as *Sphagnum cuspidatum*, *S. denticulatum*, and *Cladopodiella fluitans*. In more degraded scenarios or where high seasonal water fluctuation occurs, the pools contain open water and/or algae. Tear pools are

found on bogs where internal tensions, due to mass movement of peat, has taken place within the high bog and has caused the development of elongated pools. These are frequently found on western bogs and may be natural or anthropogenic in origin.

Hollows

These are shallow depressions (less than 5cm deep) on the bog surface where surface water collects, or where the water table reaches or lies just above ground level, depending on seasonal conditions. They are often filled with *Sphagnum* species such as *S. papillosum* and *S. cuspidatum*. They take many forms but are often eye shaped. Marginal hollows tend to be elongated as they are focused points for surface water run-off. They are often dominated by *Narthecium ossifragum*.

Lawns

These are shallow hollows or flat areas where one species dominates to form a lawn. This is frequently a *Sphagnum* species, such as *Sphagnum* magellanicum, or *S. papillosum* which can completely fill in a hollow to form a small lawn.

Flats

These are more or less flat areas which are intermediate between hollow and hummock communities. They tend to be drier than the above situations.

Hummocks

These are mounds on the bog surface which can range from a few centimetres to more than one metre in height. They are usually composed mainly of *Sphagnum* species, such as *Sphagnum magellanicum*, *S. capillifolium*, *S. austinii* and *S. fuscum* but other bryophyte species such as *Hypnum jutlandicum* and *Leucobryum glaucum* are also important, especially as the hummock grows taller and becomes drier. *Calluna vulgaris* is another important element, as it flourishes where the water table is not at surface level (Kelly & Schouten 2002).

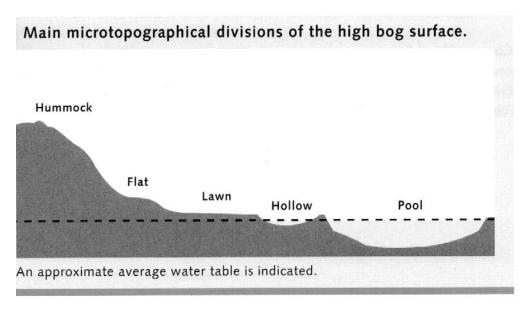


Figure 1 Raised bog microtopographical divisions on the high bog surface (reproduced from Kelly & Schouten 2002).

1.1.2 Typical Flora of Irish Raised Bogs

Raised bogs are characterised by a distinctive vegetation dominated by a variety of mosses (e.g. *Sphagnum spp., Hypnum spp., Racomitrium spp.*), sedges and grass-like species (e.g. *Eriophorum spp., Rhynchospora spp., Narthecium ossifragum, Molinia caerulea* and *Carex spp.*), and dwarf shrubs (e.g. *Calluna vulgaris, Erica tetralix, Vaccinium spp.* and *Empetrum nigrum*). In addition to these groups, a number of other species characterise raised bogs including carnivorous plants (e.g. *Drosera spp., Utricularia spp.*), lichens of both the bog surface and epiphytes on the stems of dwarf shrubs and the occasional trees on bogs (e.g. *Cladonia spp., Usnea spp.*). Herbaceous plants are not a significant element on raised bogs and include a few commonly occurring species such as *Menyanthes trifoliata, Pedicularis sylvatica*, and *Potentilla erecta* (Cross 1990).

Drier areas and hummocks usually support *Calluna vulgaris, Eriophorum vaginatum, Trichophorum germanicum, Erica tetralix,* lichens (*Cladonia* spp.), bog mosses (*Sphagnum capillifolium, S. austinii, S. fuscum, S. papillosum*) and other mosses (*Dicranum scoparium, Leucobryum glaucum*). Wet hollow areas and pools are characterised by *Eriophorum angustifolium, Rhynchospora alba, Narthecium ossifragum, Drosera* spp., *Menyanthes trifoliata,* bladderworts (*Utricularia* spp.) and bog mosses (*Sphagnum cuspidatum, S. denticulatum and S. magellanicum*).

A list of flora species that are regarded as being typical of ARB habitat in Ireland is presented in Table 1. A number of these typical species would have a restricted distribution and do not occur throughout the range of the habitat in Ireland (see above), therefore only a subset of these species would be expected to be present on any individual bog.

Table 1 Flora species typically associated with active raised bog in Ireland (after NPWS 2013). Species list is based on vegetation communities defined by Kelly (1993) and Kelly & Schouten (2002).

Common name	Scientific Name
Bog rosemary	Andromeda polifolia
Bog bead moss	Aulacomnium palustre
Bristly Swan-neck moss*	Campylopus atrovirens*
Lichen	Cladonia ciliata
Lichen	Cladonia portentosa
Long leaved sundew	Drosera anglica
Intermediate leaved sundew*	Drosera intermedia*
Round leaved sundew	Drosera rotundifolia
Common cotton grass	Eriophorum angustifolium
Hare's tail cotton grass	Eriophorum vaginatum
Large white moss	Leucobryum glaucum
Bogbean	Menyanthes trifoliata
Bog asphodel	Narthecium ossifragum
Purple spoonwort*	Pleurozia purpurea*
Woolly fringe moss*	Racomitrium lanuginosum*
White beak-sedge	Rhynchospora alba
Austin's bog moss	Sphagnum austinii
Red bog moss	Sphagnum capillifolium
Feathery bog moss	Sphagnum cuspidatum
Cow-horn bog moss*	Sphagnum denticulatum*
Rusty bog moss	Sphagnum fuscum
Magellanic bog moss	Sphagnum magellanicum
Papillose bog moss	Sphagnum papillosum
Golden bog moss*	Sphagnum pulchrum*
Lustrous bog moss	Sphagnum subnitens
Bladderwort	Utricularia minor
Cranberry	Vaccinium oxycoccos

Notes: * Species more typical of western raised bog sites.

1.1.3 Typical Fauna of Irish Raised Bogs

Raised bogs are extremely nutrient poor ecosystems. Acidic, waterlogged and exposed conditions make them an unattractive habitat for animal life. As a consequence they are relatively poor both in terms of species diversity and population densities. Many species are opportunists, vagrant or temporary rather than specialists, but nonetheless may have an important impact on the ecosystem through nutrient imports and exports or other interactions (Cross 1990). A list of fauna species that would be typically associated with raised bog habitat in Ireland is presented in Table 2. The species listed are not confined to ARB and most, if not all, will other areas of the bog and surrounding habitats.

Raised bog is unsuitable habitat for many vertebrates due to the lack of available foraging and suitable breeding places. The Irish hare is the only mammal commonly occurring. The common frog is the most common vertebrate predator.

Although 18 species of birds have been reported breeding on raised bogs (Wilson 1990) many of these species utilise the bog as a nesting habitat only. They are dependent on other neighbouring habitats such as open water bodies, callows and wet grassland particularly for feeding. Just a few species of bird, including meadow pipit (*Anthus pratensis*), skylark (*Alauda arvensis*) and curlew (*Numenius arquata*) complete their full breeding cycle on the bog and the first two species are the commonest species occurring (Bracken *et al.* 2008).

Red grouse (*Lagopus lagopus*) must also be included as a typical bog species, occurring year round as a resident. Red grouse and curlew have declined significantly on raised bogs in recent times. BirdWatch Ireland have published an Action Plan for Raised Bog Birds in Ireland which lists 13 species of conservation concern that are associated with Raised Bogs (O'Connell 2011). A recent review of birds of conservation concern in Ireland has since added meadow pipit (*Anthus pratensis*) to the red (most endangered) list of Birds of Conservation Concern in Ireland (BoCCI) (Colhoun & Cummins 2013).

Our knowledge of the invertebrate assemblages associated with Irish raised bogs remains incomplete (particularly micro-invertebrate species) with few studies undertaken (Reynolds 1984a; Reynolds 1984b; Reynolds 1985; De Leeuw 1986; O Connor *et al.* 2001; Crushell *et al.* 2008; Hannigan & Kelly-Quinn 2011; Wisdom & Bolger 2011, Nolan 2013). Van Duinen (2013) highlights the importance of structural diversity at various spatial scales (e.g. microscale of hummock hollow topography to macro-scale which would include the landscape setting of the bog, see Schouten (2002)) as a prerequisite for hosting the full species diversity of raised bog landscapes.

A recent study of Lepidoptera associated with raised bogs identified two species that appear to be characteristic of higher quality raised bog habitat, namely bordered grey (*Selidosema brunnearia* (Villers, 1789)) and light knot grass (*Acronicta menyanthidis* (Esper, 1789)) (Ciara Flynn pers. comm.).

Recent research on spiders has revealed that a number of species are known to occur in Ireland only on raised bog habitats, all of which are considered local/uncommon or rare across Europe (Myles Nolan pers. comm.). Five of these species that can be considered useful indicators of ARB include: *Glyphesis cottonae* (La Touche 1945), *Walckenaeria alticeps* (Denis 1952), *Satilatlas britteni* (Jackson 1913), *Pirata piscatorius* (Clerck 1757), and *Minicia marginella* (Wider 1834) (Myles Nolan pers. comm.).

The information currently available on other invertebrate groups of peatland systems in Ireland is not sufficient to allow a determination of many species that are typically associated with or may be characteristic of higher quality ARB. A selection of invertebrate species and species groups that are known to be typically associated with raised bogs are presented in Table 2.

Table 2 Fauna species typically associated with raised bog ecosystems in Ireland (after O'Connell 1987; Cross 1990; Renou-Wilson *et al.* 2011; Bracken & Smiddy 2012).

Common name	Scientific name
Mammal species	
Irish hare	Lepus timidus hibernicus
Otter	Lutra lutra
Pygmy shrew	Sorex minutes
Fox	Vulpes vulpes
Bird species	,
Skylark	Alauda arvensis
Mallard	Anas platyrhynchos
Greenland white-fronted goose	Anser albifrons flavirostris
Meadow pipit	Anthus pratensis
Hen harrier	Circus cyaneus
Cuckoo	Cuculus canorus
Merlin	Falco columbarius
Kestrel	Falco tinnunculus
Snipe	Gallinago gallinago
Red grouse	Lagopus lagopus
Curlew	Numenius arquata
Golden plover	Pluvialis apricaria
Lapwing	Vanellus vanellus
Reptiles and amphibians	
Common lizard	Lacerta vivipara
Common frog	Rana temporaria
Typical invertebrates	
Black slug	Arion ater
Large heath butterfly	Coenonympha tullia
Marsh fritillary butterfly	Euphydryas aurinia
Bog-pool spider	Dolomedes fimbriatus
Water striders	Gerris and Velia species
Oak eggar moth	Lasiocampa quercus
Four-spotted chaser dragonfly	Libellua quadrimaculata
Fox moth	Macrothylacia rubi
Ant	Myrmica ruginodis
Emperor moth	Saturnia pavonia
Great green bog grasshopper	Stethophyma grossa
Other species groups that are well	Araneae (spiders and mites)
represented on raised bogs include:	Ceratopogonidae (biting-midges)
	Chironomids (non-biting midges)
	Coleoptera (beetles)
	Collembola (springtails)
	Diptera (true flies)
	Dytiscidae (water beetles)
	Hemiptera (true bugs)
	Hymenoptera (bees, wasps, ants and sawflies)
	Lepidoptera (butterflies and moths)
	Odonta (dragonflies and damselflies)
	Orthoptera (grasshoppers)
	Syrphidae (hoverflies)
	Tipulidae (craneflies)
	Tabanidae (horseflies)

1.2 Habitats Directive Raised Bog Habitats in Ireland

Four habitat types listed on Annex I of the EU Habitats Directive are typically associated with raised bogs in Ireland, two of which are priority habitats (*):

- 7110 Active raised bogs (ARB)*
- 7120 Degraded raised bogs still capable of natural regeneration (DRB)
- 7150 Depressions on peat substrates of the Rhynchosporion
- 91D0 Bog woodland*

The interpretation manual of EU habitats gives the following description for 'active raised bogs': "Acid bogs, ombrotrophic, poor in mineral nutrients, sustained mainly by rainwater, with a water level generally higher than the surrounding water table, with perennial vegetation dominated by colourful Sphagna hummocks allowing for the growth of the bog (Erico-Sphagnetalia magellanici, Scheuchzerietalia palustris p., Utricularietalia intermediominoris p., Caricetalia fuscae p.). The term "active" must be taken to mean still supporting a significant area of vegetation that is normally peat forming, but bogs where active peat formation is temporarily at a standstill, such as after a fire or during a natural climatic cycle e.g., a period of drought, are also included." (CEC 2007).

DRB should be, according to the interpretation manual capable of regeneration to 'Active Raised Bog' in 30 years if appropriate measures are put in place (i.e. no major impacting activities are present and any necessary restoration works are implemented).

In Ireland, the identification of ARB is made at ecotope level based on the vegetation classification developed by Kelly (1993) and Kelly & Schouten (2002).

Raised bog vegetation communities are grouped into a series of community complexes and these complexes are then amalgamated into a series of ecotopes characterised by different physical characteristics using the approach outlined by Kelly & Schouten (2002).

The main ecotopes that community complexes are grouped into include:

- Central ecotope
- Sub-central ecotope
- · Active flushes and soaks
- Sub-marginal ecotope
- Marginal ecotope
- Inactive flushes
- Face-bank ecotope

Actively accumulating peat conditions occur within the sub-central and central ecotopes, which are the wettest on the bog and an indication of good quality ARB. Active flushes and soaks are also dominated by *Sphagnum* mosses and typically have wet conditions. These features are associated with ARB and contribute to the overall diversity of the habitat.

The adjacent surrounding marginal, sub-marginal, and face-bank bog areas typically have a supporting function for the central and sub-central communities but are not peat accumulating. These drier ecotopes may or may not correspond to the Annex I habitat DRB, as it depends on whether they are capable of regeneration to ARB. Other drier ecotopes recorded on the high bog that do not correspond to ARB include 'inactive flushes' which typically have a low *Sphagnum* cover.

The Annex I habitat Rhynchosporion depressions (7150) typically occurs along pool edges and on flats underlain by deep, wet and quaking peat. Typical plant species include Rhynchospora alba, Drosera anglica, Narthecium ossifragum, Sphagnum cuspidatum, S. denticulatum, S. magellanicum, S. papillosum, Menyanthes trifoliata, and Eriophorum angustifolium.

The priority Annex I habitat bog woodland is also actively peat-forming and overlaps with the ARB habitat. Such woodlands are usually dominated by *Betula pubescens* with a characteristic ground cover dominated by *Sphagnum* moss species, which often form deep carpets, and other mosses including species of *Polytrichum*. Woodland areas are occasionally found on raised bogs that have an absence of the characteristic moss layer and are not regarded as peat forming. Such areas do not correspond to the Annex I habitat.

1.2.1 Restoration of Active Raised Bog in Ireland

As already mentioned in the section 1.1, ARB is currently considered to be in unfavourable bad conservation status in Ireland. In addition, according to its definition, DRB should be capable of regeneration to ARB in a 30-year timescale. Thus, it follows that restoration measures are required in order to halt further losses and increase the area of ARB as well as to improve the condition of existing areas of the Annex I habitat.

Most of the restoration works undertaken so far in Ireland have concentrated on the high bog (e.g. Clara Bog, Mongan Bog, Sharavogue Bog and Raheenmore Bog) to prevent further losses as well as to restore areas to ARB. Nevertheless, some restoration works have also been undertaken on cutover areas such as at Ballykenny and Fisherstown Bogs and Killyconny Bog. Such work aims to do one or more of the following (depending on the bog in question): restore ARB on the high bog; reduce further ARB and DRB loss on the high bog; restore peat forming habitats (such as ARB, bog woodland, poor fen) on the cutover.

Works undertaken by the NPWS have indicated that there are significant differences, both ecological and economic, when comparing the effectiveness of works carried out on the cutover with those carried out on the high bog. Positive and significant results (i.e. expansion or development of ARB) can be achieved over a relatively short timeframe (10 years) on favourable areas of the high bog by blocking high bog drains. In contrast, a longer time period (30 years+) is required to achieve active peat formation on cutover areas, and even then the results are generally confined to smaller areas; i.e. flat areas (≤0.3% surface slope) or enclosed depressions that have sufficient water flow (minimum catchment 0.5ha) to maintain wet conditions throughout the year. A longer time period (minimum 50-100 years) is likely to be required for high quality ARB habitat (vegetation structure and species diversity) to develop on such cutover areas. In addition, costs of restoration measures on cutover areas are typically significantly higher than those on high bog areas.

1.3 Camderry Bog SAC

The SAC includes the raised bog, known as Camderry Bog and surrounding areas which include cutover bog and a small area of farmland.

The SAC has been selected for three Annex I habitats:

- [7110] Active raised bogs*
- [7120] Degraded raised bogs still capable of natural regeneration
- [7150] Depressions on peat substrates of the Rhynchosporion

Camderry Bog is part of a cluster of bogs in Co. Galway, situated approximately 12km northeast of Mountbellew and 9km south-east of Glenamaddy. It is almost entirely within the townlands of Camderry, Boggauns and Corracullin. The site comprises a relatively large raised bog that includes both areas of high bog and cutover bog. The northern and western margins of the site are bounded by the Shiven River, the eastern margin is bounded by a mineral ridge. A road runs parallel to the south-eastern margin of the site.

Much of the high bog has vegetation typical of the western raised bog type, and consists of two domes separated by a broad ridge of mineral soil. Overall the northern dome appears to be quite dry with limited areas of wet hummock/hollows. The lower southern dome contains an area of quaking bog with hummocks and tear pools. A small flushed area showing small-scale hummock/hollow development is found on the northern dome. To the east there is an extensive flush with areas of open water. Cutover bog occurs all around the margins of the high bog, apart from a semi-natural margin to the north by the Shiven River.

A forestry plantation on the north-western side of the bog was clearfelled in 2005 as part of a LIFE-funded restoration project undertaken by Coillte.

Recent land use includes peat-cutting, which has occurred in the south, east and north-west using mechanised methods. Damaging activities include drainage and extensive and frequent burning of the high bog.

1.3.1 Flora of Camderry Bog

ARB on Camderry Bog includes central and sub-central ecotope as well as active flush. During the most recent survey of the site (undertaken in 2012), central ecotope (0.2ha) was found at two locations with both in the southern lobe of the high bog. Both areas comprise high and low hummocks, hollows, lawns and pools. There is a 51-75% cover of *Sphagnum* dominated by *Sphagnum cuspidatum* in pools, *S. papillosum* and *S. capillifolium* with *S. magellanicum*, *S. subnitens* and *S. fuscum* also present. *Calluna vulgaris* and *Eriophorum vaginatum* dominate the vegetation with *Rhynchospora alba* also frequent. There is also some evidence of flushing as indicated by the presence of *Aulacomnium palustre* and *Polytrichum strictum* (Fernandez *et al.* 2014a, b).

Sub-central ecotope (4.6ha) was found in three areas; one in the northern lobe and two in the southern lobe of Camderry Bog. Two community complex types were recorded, one is characterised by low hummocks, hollows, flats and pools. Within this complex pool cover is 11-25% and *Sphagnum* cover 34-50%. *Sphagnum* cuspidatum dominate in pools although cover was patchy in places while hummocks of *S. capillifolium* dominate the inter-pool *Sphagna* layer. *Calluna* vulgaris, *Eriophorum* vaginatum, and *Eriophorum* angustifolium dominated the vegetation. The other community complex is restricted to areas in the southern lobe of the high bog. The pool and *Sphagnum* cover are similar to the previous complex, but here *Narthecium* ossifragum flats dominates, interspersed with *Calluna* vulgaris-dominated hummocks and shallow pools. Many of the pools are algal with *Rhynchospora* alba and some *Eriophorum* angustifolium, but most pools have a good cover

of Sphagnum cuspidatum and/or S. papillosum with S. papillosum lawns at the edges (Fernandez et al. 2014a, b).

One active flush was recorded at Camderry, located in the north of the northern lobe. The flush occurs on a gentle slope and local depression and is characterised by low hummocks, hollows, flats, lawns and pools. Pool cover is 11-25% and *Sphagnum* cover 51-75%. The central depression is a quaking lawn of *Sphagnum* and may be a small soak. Here, *Sphagnum papillosum* and *S. magellanicum* are abundant and *S. fallax* and *Aulacomnium palustre* are frequent. Where the flush grades into the surrounding sub-marginal ecotope, a flushed subcentral type of vegetation with 34-50% *Sphagnum* cover, mainly *S. magellanicum*, is present. Where drier but still active, the vegetation grades into vigorous *Eriophorum vaginatum* and *Calluna vulgaris* with some *Empetrum nigrum* (Fernandez *et al.* 2014a, b).

Inactive bog at Camderry Bog includes the sub-marginal, marginal, and face bank ecotope, as well as inactive flush. Although some areas within these degraded ecotopes have a relatively well-developed raised bog flora, they are affected by water loss to varying degrees, and are usually devoid of permanent pools. The sub-marginal ecotope (100.6ha) features the most developed microtopography. Although pools are mostly absent from the high bog, they are found within the wettest sub-marginal ecotope community complex, however, these pools generally feature a low *Sphagnum* cover and are mostly covered with algae. This complex is found in an area previously classified as sub-central ecotope and there are signs of flushing/water movement to the edge of the high bog. Pools which were again mostly algal, were also recorded within another complex, found in the southern lobe and in the northern lobe. The most common sub-marginal complex at Camderry Bog was found in large areas across the entire high bog. The *Sphagnum* cover, which ranges from 11 to 25%, is composed almost entirely of hummocks of *S. capillifolium* and *S. papillosum*. *Calluna vulgaris* dominates the vegetation along with *Eriophorum vaginatum* and *Narthecium ossifragum* while *Racomitrium lanuginosum* is notably abundant in places (Fernandez *et al.* 2014a, b).

Marginal ecotope (62.6ha) is slightly drier than sub-marginal ecotope and mainly occurs as a narrow band near the margins of the high bog. The *Sphagnum* cover is even lower here than in the sub-marginal ecotope (<10%). Four marginal complexes were recorded on Camderry Bog. The best quality marginal complex is located in the east of the south-west lobe. It appears to be slightly flushed and is dominated by *Calluna vulgaris*, *Eriophorum vaginatum*, and *Myrica gale*. It is similar to sub-marginal ecotope, but is firmer underfoot and has a lower *Sphagnum* cover. The most common marginal complex is dominated by *Carex panicea*, *Calluna vulgaris*, and *Narthecium ossifragum* with prominent *Trichophorum germanicum* in places. The two other complexes are found in association with an inactive flush in the east of the northern lobe and erosion channels on the steeply sloping ridge between the northern and southern lobes. These two complexes have some very wet areas, but the *Sphagnum* cover is low and the ground firm underfoot. (Fernandez *et al*. 2014a, b).

Face bank ecotope (3.3ha) characterised by firm ground, tall *Calluna vulgaris*, poor *Sphagnum* cover and a flat microtopography forms a very narrow band along parts of the high bog margin.

The bog also features six inactive flushes. These cover an area of 9.3 ha in total and are all located in the northern lobe of the high bog. Three are dry flushes dominated by *Molinia caerulea* with a *Sphagnum* cover of <10% dominated by *S. capillifolium*. Of the other three, two are small features; one appeares to be largely based around two artificial looking pools that were dominated by *Molinia caerulea* while the other, which is adjacent to the southeastern corner of area of clearfell, is relatively wet and had a higher *Sphagnum* cover (11-25%) including small patches of *Sphagnum fallax* and *S. palustre*. This flush is dominated by *Calluna vulgaris* and *Eriophorum vaginatum* with other species recorded including

Aulacomnium palustre, Andromeda polifolia, Polytrichum commune, Empetrum nigrum, Dryopteris dilatata, Vaccinium myrtillus, and Pinus contorta. The sixth flush also has a Sphagnum cover of 11-25% and occurs in a broad linear depression that was, at least partially, formed as a result of very old turf cutting. It appeared to receive water from a large portion of the northern lobe of the high bog and is thus wet in places, but evidence suggested that the water levels here fluctuated greatly (Fernandez et al. 2014a, b).

The area of clearfell (13.4ha) on the high bogwas previously a conifer plantation, composed mainly of *Pinus contorta* with a small area of *Picea sitchensis* andwas removed in 2005. *Sphagnum* regeneration has occurred, but as yet is mainly concentrated in the blocked forestry drains with *S. cuspidatum*, *S. fallax* and *S. palustre* recorded as well as *S. capillifolium*. This vegetation is dominated mainly by *Eriophorum vaginatum* and *Calluna vulgaris* with *Molinia caerulea*, *Juncus effusus* and small amounts of *Rubus fruticosus*, *Ulex europaeus*, and *Dryopteris dilatata* also recorded as well as scattered *Betula pubescens* saplings (circa 1m tall). Many *Pinus contorta* saplings and some *Picea sitchensis* saplings up to 3m tall were present at the edges of the clearfell (Fernandez *et al.* 2014a, b).

Rhynchosporion vegetation is widespread on Camderry Bog. It is found in both ARB and DRB, but tends to be best developed and most stable in the wettest areas of ARB. In these areas, the Rhynchosporion vegetation occurs within *Sphagnum* hollows and along *Sphagnum* pool edges and on lawns. Typical plant species include *Rhynchospora alba*, *Sphagnum cuspidatum*, *S. magellanicum*, *S. papillosum*, *Drosera anglica* and *Eriophorum angustifolium*. *Rhynchospora alba* is also found within surrounding supporting high bog habitats, but always associated with wet features such as hollows and run-off channels (Fernandez *et al.* 2014a,b).

1.3.2 Fauna of Camderry Bog

The common frog (*Rana temporaria*) occurs on Camderry Bog. Two mammals have been recorded from the SAC: Irish hare (*Lepus timidus hibernicus*) has been noted on the bog and fallow deer (*Cervus dama*) uses the bog margins and cutover.

The following three notable bird species have been reported from the bog; red grouse (*Lagopus lagopus*), snipe (*Gallinago gallinago*), and curlew (*Numenius arquata*) (DEHLG 2004; Cummins *et al.* 2010).

2 Conservation objectives

A site-specific conservation objective aims to define the favourable conservation condition of a habitat or species at site level. The maintenance of habitats and species within sites at favourable condition will contribute to the maintenance of favourable conservation status of those habitats and species at a national level.

Conservation objectives for habitats are defined using attributes and targets that are based on parameters as set out in the Habitats Directive for defining favourable status, namely area, range, and structure and functions. Attributes and targets may change or become more refined as further information becomes available.

National Conservation Objectives for raised bog SACs have recently been published in the Draft National Raised Bog SAC Management Plan (DAHG 2014). The various attributes and the justification of appropriate targets used to define favourable conservation condition for ARB relevant to Camderry Bog SAC are discussed in the following sections.

2.1 Area

NPWS has commissioned a number of raised bog surveys between 1993 and the present-Kelly et al. (1995); Derwin & MacGowan (2000); Fernandez et al. (2005); Fernandez et al. (2006); Fernandez et al. (2014). Mapping from these surveys has been used to derive the area of ARB for each bog as shown in Table 3. More recent surveys have been able to employ more precise and detailed mapping techniques and more standardised ecotope descriptions. NPWS undertook a review of data from earlier surveys in 2014 taking into account these improved techniques with the aim of providing more accurate figures for ARB. This in some cases has resulted in a change in ARB area for these earlier time periods (NPWS, unpublished data).

The national SAC target for the attribute 'habitat area' has been set at 2,590ha (DAHG 2014). This target is based on the estimated area of ARB (1,940ha) and DRB (650ha) present within the SAC network in 1994 (when the Habitats Directive came into effect).

The area of ARB at Camderry Bog in 1994 is estimated to have been 8.7ha, while the area of DRB is estimated to have been 15.9ha at that time (see Table 3). Using the same approach that has been adopted in setting the national SAC target, the site-specific target for Camderry Bog would equate to 24.6ha (sum of ARB and DRB in 1994). However, in setting the site-specific target the current hydro-ecological conditions on the bog (including cutover) have been considered in order to ensure that the target being set is based on a realistic appraisal of what is achievable as set out below.

The most recent monitoring survey of the bog estimated the area of ARB to be 6.2ha (Fernandez *et al.* 2014a, b). This represents a decrease of 2.5ha (28.7%) during the period 1994-2012. An additional survey undertaken in 2004 shows that this decrease seemed to occur steadily through the 18 year period and is likely to be attributable to negative effects resulting from the many drains that are still active on Camderry Bog (Fernandez *et al.* 2014a, b).

The current extent of DRB as estimated using a recently developed hydrological modelling technique, based largely on Light Detection and Ranging (LiDAR)² data is 22.5ha (see DAHG 2014 for further details of the technique). This represents the area of the high bog, which does not currently contain ARB but has topographical conditions deemed suitable to support

² LiDAR is a remote sensing technology that measures vertical surface elevation by illuminating a target with a laser and analysing the reflected light. This provides much more detailed topographical maps than can be collected by traditional surveying techniques.

ARB (see Map 1 which shows the total area of current and modelled potential ARB). This area was refined to 15.7ha by estimating the area that could be restored by blocking drains on the high bog. This refinement was based on applying an efficacy factor (see DAHG 2014).

Based on the current assessment of the bog, it is therefore concluded that the maximum achievable target for ARB on the high bog is 21.9ha. However, it is important to note that this assumes no further decline of ARB due to impacting activities. Similarly, should the bog be significantly dependent on regional groundwater levels then any deepening of drains in the cutover or arterial drainage works of the adjacent Shiven River could further impact the potential restoration of ARB on the high bog.

Table 3 Area of ARB and DRB recorded on the high bog at Camderry Bog in 1994, 2004, and 2012 (Source: Kelly *et al.* 1995; Fernandez *et al.* 2014a, b).

1994		2004		2012	
ARB (ha)	DRB (ha)	ARB (ha)	DRB (ha)	ARB (ha)	DRB (ha)
8.7	15.9	7.4	Unknown	6.2	15.7

A recent eco-hydrological assessment of the cutover surrounding the high bog undertaken as part of the restoration planning process estimates that, by implementing appropriate management, an additional 0.7ha of ARB could be restored in this area. The long term achievable target for ARB on Camderry Bog is therefore set at 22.6ha which is 2.0ha less than the estimated area of ARB and DRB in 1994.

In conclusion, the site-specific target for the attribute habitat area is: **Restore area of active** raised bog to 22.6ha, subject to natural processes.

2.2 Range

At a national scale, range represents the geographic range that encompasses all significant ecological variations of the ARB habitat. The national SAC target for the attribute 'range' has been set as 'not less than current range subject to natural processes'.

However, range, in the form of habitat distribution, may also be important at the site level, particularly within larger SACs, including those containing a number of individual bogs (i.e. complexes). The attribute therefore under the parameter of range is 'habitat distribution'. At the local level, it is important to conserve the variability and distribution of ARB across a raised bog SAC. This will help to ensure the diversity of the habitat is maintained while lessening the impact of localised damaging activities such as fire.

The conservation of ARB within Camderry Bog as set out in Section 2.1 above will contribute to safeguarding the national range of the habitat. The ARB habitat at Camderry includes central and sub-central ecotopes, as well as active flush. A map showing the most recent distribution of ecotopes throughout Camderry Bog is presented in Map 2.

The site-specific target for the attribute habitat distribution is: **Restore the distribution and variability of active raised bog across the SAC.**

2.3 Structure and functions

Structure and functions relates to the physical components of a habitat ("structure") and the ecological processes that drive it ("functions"). For ARB these include attributes such as the hydrological regime, water quality, habitat quality, species occurrence, elements of local distinctiveness, marginal habitats, negative physical indicators, and negative species

occurrence. As several of these attributes are inter-connected, they are all included in order to better define habitat quality in a meaningful way. In some cases, attribute targets are not quantified; however, as more detailed information becomes available (for example through further research), more measurable site-specific targets may be developed. Structure and functions attributes are expanded on in the sections below.

2.3.1 High bog area

On individual raised bogs adequate high bog is required to support the development and maintenance of ARB. Raised bog habitat that is classified as neither ARB nor DRB is still important particularly as a supporting habitat for those listed in Annex I of the Habitats Directive. It is an essential part of the hydrological unit which supports the ARB and DRB habitats. High bog is of value in its own right as a refuge for species characteristic of drier bog conditions as well as for providing a transitional zone between the Annex I habitats of the high bog and surrounding areas. Additional values for the maintenance of high bog include the preservation of its record of past environmental conditions and carbon storage. The area of high bog in the entire SAC network in 1994 was 10,740ha. The corresponding area in 2012 is 10,515ha – indicating that there has been a 225ha loss of high bog since 1994.

The national target for the attribute 'high bog' habitat is to ensure no decline in extent of high bog to support the development and maintenance of ARB.

The area of high bog within Camderry Bog SAC in 1994 was mapped as 195.3ha, while the corresponding area in 2012 is 193.2ha (based on interpretation of LiDAR and aerial photography flown in 2012), representing a loss of 2.1ha of high bog (DAHG 2014). The extent of high bog within the SAC in 2012 is illustrated on Map 1.

The site-specific target for the attribute high bog is: **No decline in extent of high bog** necessary to support the development and maintenance of active raised bog.

2.3.2 Hydrological regime: water levels

Hydrological processes are key drivers of raised bog ecology. The different raised bog communities, assemblages and species are affected by various hydrological attributes. For ARB, mean water levels need to be near or above the surface of bog lawns for most of the year. Seasonal fluctuations should not exceed 20cm, and water level should be within 10cm of the surface, except for very short periods of time (Kelly & Schouten 2002). Gentle slopes that limit intermittent lateral losses of water (through surface runoff) and encourage sustained waterlogging are the most favourable to achieve these conditions. Such conditions may be maintained on steeper slopes in areas of focused flow (flushes).

The traditional view of water flowing across the bog laterally has been recently refined to also consider that water flows vertically through peat into the underlying substrate. Water loss, by this route, depends on the permeability of the material through which the water must flow and the difference in head (water level elevation) in the bog and underlying mineral substrate; larger differences encountered in higher permeability materials will result in greater losses. Although the proportion of water lost in this manner may be small, the sustained loss during prolonged dry periods may be sufficient to impact bog ecotopes. Drains extending into the mineral substrate in marginal areas surrounding the bog can lead to an increased gradient between the head in the peat and the head in the underlying substrate resulting in increased vertical water losses from the bog.

The most recent description of drainage at Camderry Bog is presented in Fernandez et al. (2014a, b) who reported that at least 5.5 km of drains were impacting upon the raised bog

habitats. A total of 2.9km of these drains were reported as reduced functional as they were starting to infill with vegetation while 2.6km were reported as functional. There were also 4.4km of unblocked drains in-filled with vegetation; it is not known whether these drains are impacting on the hydrology of the bog. Both the functional and reduced functional drains are considered to be impacting on high bog habitats and will continue to do so until they become completely non-functional. There are also a number of blocked drains which are all considered to have reduced functionality.

There are cutover drains associated with former peat cutting present along the entire south, east, and north-west of the northern lobe and along the west of the southern lobe. These drains continue draining the high bog and impacting on high bog habitats. The main arterial drain to Camderry Bog is the Shiven River, located close to the south-west boundary.

Much of the knowledge regarding the hydrological requirements of raised bog communities in Ireland stems from the extensive ecological and hydrological work undertaken on Clara Bog since the early 1990s. The only available hydrological study for Camderry Bog is the work carried out by Kelly *et al.* (1995). The hydrochemistry survey of the cutover drains identified relatively low electrical conductivity (EC) values, indicating very little groundwater contributions to the drains at the time of the survey. EC values of $<80\mu$ S/cm are typical for rainwater and also bog water due to the low reactivity of the peat. Typical values in the south of the bog were between 90-110 μ S/cm and in the west 90 μ S/cm. At the southern lobe of bog, the river had a higher EC value of 378 μ S/cm suggesting much greater contributions of mineralised groundwater.

Kelly et al. (1995) noted that the tributary of the Shiven River, which runs close to the bog at its north, north-west, and south-west edges appeared to have been drained. Maintenance of this channel is likely to result in lowering of the regional groundwater table. Fernandez et al. (2014a, b) considered bog margin drainage, including arterial drainage and river dredging works, to have a medium importance / impact on high bog habitats. This drainage has the potential to result in impacts on regional groundwater heads, which could result in subsidence on the surface of Camderry Bog. The risk of subsidence depends on the permeability of the underlying substrate, which will influence the extent of impacts from changes to groundwater head.

Geological mapping shows the area as directly underlain by Visean limestone, which is typically classed as a regionally important aquifer as it is subject to Karstification (conduit). Subsoil mapping shows that the bog is surrounded by sandstone till to the east and south, and limestone till to the north and west. There is also alluvium to the north and south of the bog associated with river channels. The presence of highly permeable bedrock and potentially permeable substrate underlying the peat suggests that a decline in groundwater head has a high risk of impacting high bog habitats at Camderry Bog.

The site-specific target for the attribute hydrological regime – water levels is: **Restore** appropriate water levels throughout the site.

2.3.3 Hydrological regime: flow patterns

As outlined above, ARB depends on water levels being near or above the surface of bog lawns for most of the year. Long and gentle slopes are the most favourable to achieve these conditions. Changes to flow directions due to subsidence of bogs can radically change water regimes and cause drying out of high quality ARB areas and soak systems.

A map illustrating the slopes and drainage patterns on Camderry Bog based on a digital elevation model generated from LiDAR imagery flown in 2012 is presented in Map 3.

It is evident that there has been subsidence on the surface of the bog as steep slopes (>1%)

extend into several sections of the bog, particularly in areas where extensive turf-cutting has been carried out. Flow patterns of the high bog are focused towards the margins suggesting that there has been differential subsidence on the high bog surface. This is particularly apparent along the eastern side of the bog where flow patterns can be seen converging towards the functional drains and areas where extensive peat cutting has taken place. Subsidence is evident towards the south of the bog as indicated by the steep slopes along the western margin and areas of focused flow. In this area there are intact areas of ARB; however, should further subsidence occur around the margins there is a risk that Camderry Bog would no longer be able to support ARB in this area.

The site-specific target for the attribute hydrological regime – flow patterns is: **Restore**, where possible, appropriate high bog topography, flow directions and slopes.

2.3.4 Transitional areas between high bog and surrounding mineral soils (includes cutover areas)

Transitional zones between raised bogs and surrounding mineral soils are typically cutover bog and drained lagg zones. The maintenance / restoration of these areas will help to maintain hydrological integrity of ARB and DRB, and support a diversity of other wetland habitats (e.g. wet woodland, swamp and fen) as well as species that they sustain. In some cases, these areas may assist in reducing further losses of ARB / DRB on the high bog and in time could develop into active peat forming habitats (including ARB - see Section 2.1 above). These transitional zones, once restored, can provide ecosystem services through flood attenuation and water purification to downstream areas and potentially increase the carbon storage / sink function of the bog. The estimated extent of such transitional areas within the SAC network is 3,000ha (DAHG 2014). The national target for these transitional areas is to maintain / restore semi-natural habitats with high water levels around as much of the bog margins as necessary.

The transitional areas at Camderry Bog include a range of different habitat types (e.g. wet grassland, cutover bog, scrub etc.). The total area of cutover bog within the Camderry Bog SAC is estimated to be circa 90ha. The development of habitats within cutover areas depends on a number of factors including prevailing land-use, topography, up-welling regional groundwater, and drainage.

Kelly et al. (1995) described the cutover areas of Camderry Bog to be mainly colonised by *Ulex europaeus* and *Betula pubescens* with areas of *Molinia caerulea*. At the southern edge of the southern lobe there is a *Crataegus monogyna* hedgerow very close to the bog. Some of the drains in the vicinity of a flush contain enrichment species such as *Potamogeton polygonifolius* and *Carex rostrata*.

The Shiven River lies immediately alongside the most northerly part of Camderry Bog. Kelly et al. (1995) recorded a steep gradational slope from the bog to the river. Natural erosion channels leave the bog and enter the narrow river which is only 15cm deep and 0.65m wide. Crataegus monogyna, Salix sp., and Calluna vulgaris grow along the banks of the Shiven river here as well as typical river bank species such as Filipendula ulmaria, Urtica dioica, and Rumex sp. To the west of the southern lobe old cutover is dominated by Calluna vulgaris with Juncus effusus, Eriophorum angustifolium and Molinia caerulea. Betula pubescens and Salix spp. also occur, mainly concentrated along the drains. Along the river to the southwest, Calluna vulgaris and Juncus effusus dominate. To the south there is some Sphagnum regeneration in the cutover (Kelly et al. 1995).

Kelly et al. (1995) also recorded that along the south-west margins of Camderry Bog, the bog edge is 5m from the Shiven River separated by flush vegetation. The peat is very thin along all the edges in the southern lobe. The drumlin that separates this lobe from the main bog

consists of silt till and is overgrown with *Ulex europaeus*. The facebanks here are a product of hand-cutting and are overgrown with *Calluna vulgaris*. Few drains run along this side since generally the bog recharges groundwater here.

In 2004 Fernandez et al. (2005) noted in their survey that there are several patches of good quality regenerating cutover around the high bog. Old cutover that had been abandoned for a long time occurs to the south-east of the southern lobe. These areas are dominated by Eriophorum vaginatum and other graminoids and have over 50% Sphagnum cover with S. magellanicum, S. papillosum, S. capillifolium and S. cuspidatum. On the eastern side of the northern lobe in the cutover Fernandez et al. (2005) noted old drains in-filling with Sphagnum cuspidatum, S. magellanicum and S. papillosum as well as Eriophorum angustifolium and Rhynchospora alba.

The site-specific target for the attribute transitional areas is: **Restore adequate transitional** areas to support / protect active raised bog and the services it provides.

2.3.5 Vegetation quality: central ecotope, active flush, soaks, bog woodland

A diverse good quality microtopography on raised bogs consists of *Sphagnum* dominated pools, hollows, lawns and hummocks, which support the highest diversity of species including hummock indicators: *Sphagnum fuscum* and *S. austinii*; pool indicators: *S. cuspidatum*, *S. denticulatum*, and indicators of lack of burning events e.g. some lichen species (*Cladonia* spp.) (Cross 1990).

The national target for the attribute vegetation quality has been set as "to maintain / restore sufficient high quality bog vegetation (i.e. central ecotope and / or flushes / soaks). At least 50% of ARB habitat should be central ecotope and / or flush / soaks." Bog woodland is also regarded as a desirable variant of ARB as it adds species and structural diversity to the habitat and therefore, where relevant, also contributes to the 50% target at site level.

A summary description of the vegetation of Camderry Bog is presented in Section 1.3.1 above. The vegetation and habitats of the bog have been described in more detail by Kelly *et al.* (1995) and Fernandez *et al.* (2005 & 2014a, b).

It is worth noting here that Fernandez et al. (2014a, b) described an active peat forming flush on a gentle slope in the north of the bog in which there is a local depression that may be a small soak. In this 'soak' area there is a quaking lawn of *Sphagnum* including *S. papillosum*, *S. magellanicum* and *S. fallax* with an overall *Sphagnum* cover of 51-75% while elsewhere within the flush the *Sphagnum* cover is 34-50%.

The extent of the different ecotopes that correspond with ARB based on the most recent surveys is presented in Table 4 and on Map 2. It can be seen that the proportion of ARB that comprises central ecotope and active flush (including possible soak) is currently 24.2%. Comparing this to results of surveys undertaken in 2004 indicates that although the total area of ARB decreased across the site, the proportion of each of the component ecotopes remained relatively static.

The target for this attribute is 11.3ha of central ecotope and active flush (50% of ARB target area (22.6ha)). This requires an increase of 5.1ha from the current area of central ecotope and flush (including possible soak).

Table 4 Extent of ecotopes classified as ARB in 2004 and 2012 (modified from Fernandez *et al.* 2014a, b).

Ecotope	2004		2012	
	ha	% of total ARB	ha	% of total ARB
Sub-central ecotope	5.7	77.0	4.7	75.8
Central ecotope	0.4	5.4	0.2	3.2
Soaks / active flush	1.3	17.6	1.3	21.0
Total ARB	7.4		6.2	

The site-specific target for the attribute vegetation quality is: **Restore 11.9ha of central ecotope/active flush/soaks/bog woodland as appropriate.**

2.3.6 Vegetation quality: microtopographical features

The characteristic microtopographical features of raised bogs are described in Section 1.1.1 above.

Hummock and hollow microtopography is well developed on Camderry Bog (Kelly et al. 1995; Fernandez et al. 2014a, b).

The site-specific target for the attribute microtopographical features is: **Restore adequate** cover of high quality microtopographical features.

2.3.7 Vegetation quality: bog moss (Sphagnum) species

Bog mosses, which have unique properties, are the principal component of peat, and are largely responsible for the typical microtopographical features as described in Section 2.3.6 above.

The vegetation of a typical raised bog that is still hydrologically intact is characterised by the dominance of several species of Sphagna and dwarf ericoid shrubs. The most abundant species are *Sphagnum capillifolium*, *S. austinii* and *S. papillosum* which form hummocks or low ridges. *Sphagnum fuscum* may also form hummocks (Cross 1990). On the flats *Sphagnum magellanicum*, *S. papillosum*, *S. tenellum*, and *S. subnitens* are the key species. *Sphagnum pulchrum* may also be dominant in flats on western raised bogs. In permanently waterlogged hollows *Sphagnum cuspidatum* and *S. denticulatum* (western bogs) occur. *Sphagnum fallax* is common where there is slight flushing (Cross 1990). The most commonly occurring *Sphagnum* moss species that occur on raised bogs in Ireland are presented in Table 5 along with a summary of their ecology and typical contribution to peat formation.

Kelly et al. (1995) and Fernandez et al. (2014a, b) provide information on the occurrence of *Sphagnum* species throughout Camderry Bog.

Table 5 *Sphagnum* species typically associated with raised bog ecosystems in Ireland. Ecology as described by Laine *et al.* (2009) with minor modifications.

Species	Ecology	Peat forming capacity
Sphagnum austinii	hagnum austinii Hummock species	
Sphagnum capillifolium	Forms small hummocks and carpets	Moderate
Sphagnum cuspidatum	Pool and hollow species	Low
Sphagnum denticulatum	Pool and hollow species	Low
Sphagnum fallax Occurs in lawns and carpets, shade tolerant. Indicative of some nutrient enrichment (soaks and active flushes)		Low
Sphagnum fuscum Forms dense low and wide, and occasionally high hummocks		High
Sphagnum magellanicum	Lawn species forming carpets and low hummocks	Moderate
Sphagnum palustre	Forms hummocks and dense carpets, often in shaded conditions. Indicative of nutrient enrichment (soaks and active flushes)	Low
Sphagnum papillosum	Lawn , hollow, and low hummock species	Moderate
Sphagnum pulchrum Grows in lawns and hollows, more typical of western bogs		Moderate
Sphagnum squarrosum Forms carpets and small mounds. Indicative of nutrient enrichment (soaks and active flushes)		Low
Sphagnum subnitens	Occurs as individual shoots or small cushions and lawns. Tolerant of minerotrophic conditions	Moderate
Sphagnum tenellum	Occurs as single shoots or weak cushions, typically in disturbed patches of the bog surface	Low

The site-specific target for the attribute bog moss (*Sphagnum*) species is: **Restore adequate** cover of bog moss (*Sphagnum*) species to ensure peat-forming capacity.

2.3.8 Typical ARB species: flora

Camderry Bog supports the full complement of plant species typically associated with a true midland raised bog (see Section 1.1.2 above).

The key typical species that are indicative of high quality raised bog include *Sphagnum fuscum* and *S. austinii* which are associated with hummocks and *S. cuspidatum* and *S. denticulatum* which are associated with pools and hollows. All of these species were reported from Camderry Bog in 1994 (Kelly *et al.* 1995) and 2004 (Fernandez *et al.* 2005), however, the 2012 survey found a very marked decrease in the occurrence of *Sphagnum denticulatum* with it being absent from all the quadrats where it had been record in 2004 (Fernandez *et al.* 2014a, b).

The site-specific target for the attribute typical bog flora is: **Restore, where appropriate, typical active raised bog flora.**

2.3.9 Typical ARB species: fauna

As mentioned in section 1.1.3, a list of typical fauna specific to ARB has not been developed and the table contains species that use the wider raised bog habitat. This may be refined as more information becomes available.

Camderry Bog is likely to support a wide range of fauna species that are typically associated

with raised bog habitat (see Section 1.1.3 above).

The site-specific target for the attribute typical bog fauna is: **Restore, where appropriate, typical active raised bog fauna.**

2.3.10 Elements of local distinctiveness

A range of features may be associated with raised bogs which add to the scientific, historical, or conservation value of a bog. These can include geological, topographical, archaeological and hydrological features (e.g. soaks, lakes, flushes) and noteworthy species of flora and fauna (Cross 1990). Notable species of flora and fauna include those listed in the Habitats and Birds Directives, Red-listed species, and other rare or localised species. For this attribute, features that are particularly associated with ARB are relevant.

2.3.10.1 Site features

Camderry Bog features several flush systems all of which are inactive except for one. Part of this active flush is thought to be a possible small soak (Fernandez *et al.* 2014a, b).

'Soak' is a term used in European bog literature to describe an area of fen vegetation occurring within an acid bog (Gore 1983; Cross 1990; Connolly *et al.* 2002; Rydin & Jeglum 2006; Crushell 2008). The presence of fen vegetation is due to increased nutrient / mineral supply by either minerogenous water (minerotrophic) or the lateral through-flow of ombrogenous water from the surrounding bog expanse (rheotrophic) (Osvald 1949; Gore 1983).

In contrast to the non-wooded vegetation dominated by ericaceous shrubs and *Sphagnum* mosses that typifies the oceanic raised bogs, soaks are characterised by oligo-mesotrophic open-water communities, poor fen and / or bog woodland vegetation (Cross 1990).

Soaks support a relatively high diversity of species and unusual flora and fauna assemblages not typically found within raised bogs (Overbeck 1975; Kelly 1993; Reynolds 1990) thereby adding to the heterogeneity and biodiversity of the ecosystem type (Verberk 2008).

Fernandez et al. (2014a, b) described how the active peat forming flush, located in the north of Camderry Bog, occurs on a gentle slope and local depression and is characterised by low hummocks, hollows, flats, lawns and pools. Pool cover is 11-25% and *Sphagnum* cover 51-75%. The central depression is a quaking lawn of *Sphagnum*. Here, *Sphagnum papillosum* and *S. magellanicum* are abundant and *S. fallax* and *Aulacomnium palustre* are frequent. Where the flush grades into the surrounding sub-marginal ecotope, a flushed sub-central type of vegetation with 34-50% *Sphagnum* cover, mainly *S. magellanicum*, is present. Where drier but still active, the vegetation grades into vigorous *Eriophorum vaginatum* and *Calluna vulgaris* with some *Empetrum nigrum*.

Another feature of the SAC is the presence of natural margins along some of the bog's boundaries with the nearby Shiven River as described above in Section 2.3.4 above.

2.3.10.2 Rare flora

No rare flora records have been reported from Camderry Bog.

2.3.10.3 Rare fauna

As mentioned above, there is a lack of documented site-specific data relating to species that are particularly associated with ARB, including rare species.

Breeding red grouse (*Lagopus lagopus*) have been recorded on the bog (Cummins *et al.* 2010).

In conclusion, the site-specific target for the attribute elements of local distinctiveness is: Maintain features of local distinctiveness, subject to natural processes.

2.3.11 Negative physical indicators

Raised bogs that have been damaged by marginal cutting and drainage, reclamation for agriculture, forestry activities, fire, surface drainage, or the lowering of regional water tables show a range of negative physical indicators (Cross 1990). Such negative physical features of ARB include: bare peat, algae dominated pools and hollows, marginal cracks, tear patterns, subsidence features such as dry peat and / or mineral mounds / ridges emerging or expanding, and burning evidence.

Fernandez *et al.* (2014a, b) found many algal pools in the sub-central ecotope areas during their 2012 survey. This survey also found bare peat and erosion channels between the northern

Fernandez et al. (2005) noted that an area of approximately 10ha in the south-west of the northern lobe had been significantly damaged by a fire prior to their 2004 site visit.

The site-specific target for the attribute negative physical indicators is: **Negative physical features absent or insignificant.**

2.3.12 Vegetation composition: native negative indicator species

Indicators of disturbance on a raised bog include species indicative of drying out conditions such as abundant *Narthecium ossifragum* and *Trichophorum germanicum*; *Eriophorum vaginatum* forming tussocks; abundant *Sphagnum magellanicum* in pools previously dominated by species typical of very wet conditions (e.g. *Sphagnum cuspidatum*). Indicators of frequent burning events include abundant *Cladonia floerkeana* and high cover of *Carex panicea* (particularly in the true midlands raised bog type).

The site-specific target for the attribute negative indicator species is: **Native negative indicator species at insignificant levels.**

2.3.13 Vegetation composition: non-native invasive species

Non-native invasive species that can commonly occur on raised bog habitats include: *Pinus contorta, Rhododendron ponticum,* and *Sarracenia purpurea* (Cross 1990).

Non-native species are not a major issue on Camderry Bog although *Pinus contorta* saplings and *Picea sitchensis* saplings up to 3m tall were present at the edges of the clearfell in the west of the northern lobe.

Campylopus introflexus was also recorded, particularly in the east of the northern lobe close to a flush (Fernandez et al. 2014a, b).

The site-specific target for the attribute non-native invasive species: **Non-native invasive** species at insignificant levels and not more than 1% cover.

2.3.14 Air quality: nitrogen deposition

Peatlands are highly sensitive to air pollution, particularly nitrogen deposition. Reactive nitrogen from fossil fuel combustion or intensive agriculture can contaminate rain and snow, causing soil acidification, nutrient enrichment, and a decline in species that are sensitive to these conditions. There is evidence that the combined impact of elevated nitrogen deposition and a warming climate could exceed the sum of the individual stressors and lead to a dramatic decline in the biodiversity of mosses, sensitive vascular plants, and microbes,

potentially leading to catastrophic peat loss (PEATBOG project http://www.sste.mmu.ac.uk).

Air pollution can change both the species composition and the functioning of peatlands. The primary atmospheric pollutant from the Industrial Revolution to the mid 1970s was sulphur deposition, but levels have since greatly declined. Reactive nitrogen (N) deposition (primarily NO3- and NH4+), which can both acidify and eutrophy, became significantly elevated over a widespread area in the early to mid-20th century and is now the major pollutant in atmospheric deposition across most of Europe (Fowler *et al.* 2005).

Nitrogen is commonly a limiting terrestrial nutrient and in un-impacted peatlands it is tightly cycled. With long-term elevated N deposition, vegetation composition typically shifts toward species adapted to higher nutrient levels, with an overall loss of diversity (Malmer & Wallén 2005). In peatlands, field experiments with N additions within the current European range have shown significant declines in bryophyte species-richness and productivity, and shifts in composition toward vascular plants (Bobbink *et al.* 1998; Bubier *et al.* 2007). Community shifts toward more nitrophilous bryophytes in N-enriched regions such as parts of the Netherlands are also well documented (Greven 1992). In the UK, both a general survey of peatlands across the country (Smart *et al.* 2003), and a targeted study of *Calluna* moorland (Caporn *et al.* 2007) showed significant inverse relationships between levels of nitrogen deposition and species richness, with bryophytes particularly impacted. Changes in the vegetation also impact below-ground communities and biogeochemical processes.

Moderate increases in N deposition from a low level may increase *Sphagnum* and vascular plant productivity without an equal increase in decomposition rates, leading to enhanced carbon accumulation (Turunen *et al.* 2004). However, shifts in species composition from bryophytes to vascular plants may increase the production of easily-decomposable plant material, leading to higher rates of decomposition, and reduced carbon accumulation (Lamers *et al.* 2000; Bubier *et al.* 2007).

The particular sensitivity of nutrient-poor ombrotrophic peatlands to nitrogen enrichment is reflected in the low critical load threshold of between 5 and 10kg N/ha/yr for these ecosystems (Bobbink & Hettelingh 2011), a level which is exceeded over a significant portion of their range. An Irish study during the late 1990s undertaken by Aherne & Farrell (2000) concluded that total N deposition shows a strong east-west gradient, with lowest deposition in the west at 2kg N/ha/yr and highest in the east and south-east at 20kg N/ha/yr. Average N deposition over the Republic of Ireland was estimated to be approximately 12kg N/ha/yr. The study also concluded that the Critical Load Threshold for N was exceeded in at least 15% of ecosystems studied. The critical load applied to peatland ecosystems by Aherne & Farrell (2000) was 10kg N/ha/yr. This is in line with the recommendation by Bobbink & Hettelingh (2011) that the critical load should be set at the high end of the range in areas of high precipitation and at the low end of the range in areas of low precipitation assuming that Ireland represents a high precipitation area.

It is recommended in the case of Camderry Bog that the level of N deposition should not exceed the low end of the range i.e. 5kg N/ha/yr. This recommendation is based on a precautionary approach, as the evidential basis for setting a higher level is not particularly strong as alluded to by Payne (2014). Total N deposition in the vicinity of Camderry Bog as reported by Henry & Aherne (2014) is 12.1kg N/ha/yr.

The site-specific target for the attribute air quality is: Air quality surrounding bog close to natural reference conditions. The level of N deposition should not exceed 5kg N/ha/yr.

2.3.15 Water quality

Ombrotrophic peat waters found on the surface of raised bogs are characterised by low pH values (pH < 4.5) (Moore & Bellamy 1974) and also have low values of electrical conductivity. This is due to the fact that the raised bog system derives its mineral supply from precipitation, which is usually acidic and low in nutrients. Raised bog vegetation exchanges cations with protons to further reduce the pH.

Hydrochemistry varies in the areas surrounding a raised bog. Locally, conditions may be similar to the high bog due to a dominance of water originating from the bog. However, elsewhere in the marginal areas, there may be increased mineral and nutrient content of the water due to regional groundwater influences, runoff from surrounding mineral soils, and the release of nutrients through oxidation of peat resulting from reduced water levels.

The hydrochemistry of Camderry Bog has not been studied in detail. Data presented by Kelly et al. (1995) recorded relatively low EC values within marginal drains surrounding the bog. Higher EC values recorded in the Shiven River suggests a significant contribution of mineralised groundwater. The presence of a possible soak suggests increased nutrient availability at that location.

The site-specific target for the attribute water quality is: **Water quality on the high bog and** in transitional areas close to natural reference conditions.

3 References

Aherne, J., & Farrell, E.P. (2000) Final Report: Determination and mapping of critical loads for sulphur and nitrogen and critical levels for ozone in Ireland. Environmental Protection Agency, Dublin, 212pp.

Bobbink, R. & Hettelingh, J.P. (2011) Review and revision of empirical critical loads and dose-response relationships. Proceedings of an expert workshop, Noordwijkerhout, 23-25 June 2010. RIVM report 680359002, Coordination Centre for Effects, National Institute for Public Health and the Environment (RIVM).

Bobbink, R., Hornung, M. & Roelofs, J.G.M. (1998) The effects of air—borne nitrogen pollutants on species diversity and semi—natural European vegetation. Journal of Ecology 86: 717–738.

Bracken, F. & Smiddy, P. (2012) Lowland bogs, fens and reedswamps, pp. 73-89. In: Nairn, R., and O'Halloran, J. (eds.) Bird Habitats in Ireland. The Collins Press, Cork.

Bracken, F., McMahon, B. & Whelan, J. (2008) Breeding bird populations of Irish Peatlands: capsule peatlands are very important habitats for birds despite low species diversity. Bird Study 55 (2): 169-178.

Bubier, J., Moore, T. & Bledzki, L.A. (2007) Effects of nutrient addition on vegetation and carbon cycling in an ombrotrophic bog. Global Change Biology 13: 1168–1186.

Caporn, S.J.M., Edmondson, J., Carroll, J.A., Pilkington, M. & Ray, N. (2007) Long-term impacts of enhanced and reduced nitrogen deposition on semi-natural vegetation. Report to Defra. Terrestrial Umbrella. Work Package 2: Impacts, Recovery and Processes. Task 4. Defra London.

CEC (2007) Interpretation manual of European Union Habitats. Version EUR 27. European Commission, DG Environment, Brussels. Nature and Biodiversity.

Colhoun, K. & Cummins, S. (2013) Birds of Conservation Concern in Ireland 2014–2019. Irish Birds 9: 523-544.

Cross, J. (1990) The Raised Bogs of Ireland, their ecology, status and conservation. Report to the Minister of State at the Department of Finance. The Stationery Office, Dublin.

Crushell, P.H. (2008) Soak Systems of an Irish Raised Bog: a multidisciplinary study of their origin, ecology, conservation and restoration. PhD thesis, Wageningen University, with a summary in Dutch and Irish.

Crushell, P.H., Schouten, M.G.C., Robroak, B.J.M. & van Duinen, G-J. (2008) The contribution of soak lakes to macroinvertebrate diversity of raised bogs in Ireland. In: Crushell, P.H. (2008). Soak Systems of an Irish Raised Bog: a multidisciplinary study of their origin, ecology, conservation and restoration. PhD thesis, Wageningen University, with a summary in Dutch and Irish.

Cummins, S., Bleasdale, A., Douglas, C., Newton, S., O'Halloran, J. & Wilson, H.J. (2010) The status of Red Grouse in Ireland and the effects of land use, habitat and habitat quality on their distribution. Irish Wildlife Manuals, No. 50. National Parks and Wildlife Service, Department of the Environment, Heritage and Local Government, Dublin, Ireland.

DAHG (2014) National Raised Bog SAC Management Plan. Draft for Consultation. Main report and appendices. Department of Arts, Heritage and the Gaeltacht.

DEHLG (2004) National Parks and Wildlife Service Conservation Plan for 2004-2009. Draft 1. Camderry Bog cSAC: Site Code 2347, County Galway.

De Leeuw, J.P.M. (1986) Een onderzoek naar het voorkomen en de verspreiding van aquatische macro- en mirofauna in de Ierse hoogvenen. Deel 1: Macrofauna. Aquatische Oecologie, Katholieke Universiteit Nijmegen, Nijmegen, The Netherlands.

Derwin, J. & MacGowan, F. (2000) Raised bog restoration project: a continuation of the investigation into the conservation and restoration of selected raised bog sites in Ireland. Unpublished report, Dúchas the Heritage Service, Dublin.

Fernandez Valverde, F., Fanning, M., McCorry, M. & Crowley, W. (2005) Raised bog monitoring project 2004-2005. Document 3: Site Reports and Maps Volume 1-5. Unpublished Report. National Parks and Wildlife Service, Dublin.

Fernandez, F., MacGowan, F., Crowley, W., Farrell, M., Croal, Y., Fanning, M. & McKee M. (2006) Assessment of the impacts of turf cutting on designated raised bogs 2003-06. Unpublished report, National Parks and Wildlife Service, Department of Environment, Heritage and Local Government, Dublin.

Fernandez, F., Connolly K., Crowley W., Denyer J., Duff K. & Smith G. (2014a) Raised bog monitoring and assessment survey 2013. Irish Wildlife Manuals, No. 81. National Parks and Wildlife Service, Department of Arts, Heritage and Gaeltacht, Dublin, Ireland.

Fernandez, F., Connolly, K., Crowley, W., Denyer J., Duff, K. & Smith, G. (2014b) Raised bog monitoring and assessment survey 2013 – Camderry Bog, Co. Galway – site report. National Parks and Wildlife Service, Department of Arts, Heritage and Gaeltacht, Dublin, Ireland.

Fossitt, J. (2000) A Guide to Habitats in Ireland. The Heritage Council, Ireland.

Fowler, D., Smith, R.I., Muller, J.B.A., Hayman, G. & Vincent, K.J. (2005) Changes in the atmospheric deposition of acidifying compounds in the UK between 1986 and 2001. Environmental Pollution, 137: 15-25.

Gore, A.J.P. (ed.) (1983) Ecosystems of the world 4A. Mires: Swamp, bog, fen and moor. General studies. Elsevier Scientific Publishing Company, Amsterdam.

Greven, H.C. (1992) Changes in the moss flora of the Netherlands. Biological Conservation 59: 133-137.

Hannigan, E., & Kelly-Quinn, M. (2011) Chapter 2.6 - Aquatic macro-invertebrate diversity. pp. 140-157 In: Renou-Wilson, F. (ed.) BOGLAND: Sustainable Management of Peatlands in Ireland. Environmental Protection Agency, Wexford.

Henry, J. & Aherne, J. (2014) Nitrogen deposition and exceedance of critical loads for nutrient nitrogen in Irish grasslands. Science of the Total Environment 470–471: 216–223.

Kelly, L. & Schouten, M.G.C. (2002) Vegetation. In: Schouten, M.G.C. (ed.), Conservation and restoration of raised bogs: geological, hydrological and ecological Studies. Dúchas — The Heritage Service of the Department of the Environment and Local Government, Ireland; Staatsbosbeheer, the Netherlands; Geological Survey of Ireland, Dublin. pp. 110-169.

Kelly, L., Doak, M. & Dromey, M. (1995) Raised Bog Restoration Project: An Investigation into the Conservation and Restoration of Selected Raised Bog Sites in Ireland. Part 1 Summary Reports. National Parks & Wildlife Service, Department of Environment, Heritage and Local Government, Dublin.

Kelly, M.L. (1993) Hydrology, hydrochemistry and vegetation of two raised bogs in county Offaly. PhD thesis, Trinity College Dublin.

Laine, J., Harju, P., Timonen, T., Laine, A., Tuittila, E.S, Minkkinen, K. & Vasander, H. (2009) The Intricate beauty of Sphagnum mosses - A Finnish guide to identification. University of Helsinki Department of Forest Ecology Publications, 39: 1–190.

Lamers, L. P. M., Bobbink, R. & Roelofs, J. G. M. (2000) Natural nitrogen filter fails in polluted raised bogs. Global Change Biology, 6: 583–586.

Malmer, N. & Wallén, B. (2005) Nitrogen and phosphorus in mire plants: variation during 50 years in relation to supply rate and vegetation type. Oikos, 109: 539–554.

Moore, P.D. & Bellamy, D.J. (1974) Peatlands. Elek Science. London.

NPWS (2008) The Status of EU Protected Habitats and Species in Ireland. National Parks and Wildlife Service, Ireland.

NPWS (2013) *The Status of EU Protected Habitats and Species in Ireland.* Version 1.0. Unpublished Report, National Parks and Wildlife Services. Department of Arts, Heritage and the Gaeltacht, Dublin, Ireland.

NPWS (2014). National Peatland Strategy - Draft for Consultation. NPWS, Department of Arts, Heritage and the Gaeltacht, Dublin, Ireland.

Nolan, M. (2013) Spiders (Araneae) of Irish raised bogs: Clara bog, Co. Offaly and Carrowbehy bog, Co. Roscommon. Bulletin of the Irish Biogeographical Society 37: 172-203.

O'Connell C. (ed.) (1987) The IPCC Guide to Irish Peatlands. Irish Peatland Conservation Council, Dublin.

O'Connell, P. (2011) Action Plan for Raised Bog Birds in Ireland 2011-2020. BirdWatch Ireland, Kilcoole, Co Wicklow.

O Connor, Á., Reynolds, J.D. & Kavanagh, B. (2001) Aquatic macroinvertebrate colonisation of artificial water bodies in cutaway oceanic raised bog in Ireland. In: Rochfort, L. and Daigle, J.Y. (eds.), Proceedings of the 11th International Peat Congress. pp. 742-750.

Osvald, H. (1949) Notes on the vegetation of British and Irish mosses. Acta Phytogeographica Suecica 26: 1-62.

Overbeck, F.Th. (1975) Botanisch-geologische Moorkunde. Wachholtz Verlag, Neumünster.

Payne, R.J. (2014) The exposure of British peatlands to nitrogen deposition, 1900–2030. Mires and Peat 14: Art. 4.

Renou-Wilson, F., Bolger, T., Bullock, C., Convery, F., Curry, J., Ward, S., Wilson, D. & Müller, C. (2011) BOGLAND: Sustainable Management of Peatlands in Ireland. STRIVE Report Series No.75. Prepared for the Environmental Protection Agency. pp. 181.

Reynolds, J.D. (1984a) Invertebrate survey of Irish midlands raised bogs. Bulletin of the British Ecological Society 15: 81-82.

Reynolds, J.D. (1984b) Invertebrate fauna of Irish raised bogs. Part II: Odonata, aquatic Hemiptera and Trichoptera. Bulletin of the Irish Biogeographical Society 8: 98-102.

Reynolds, J.D. (1985) Invertebrates of Lough Roe, Co. Offaly; a rare and endangered bogland habitat. Bulletin of the Irish Biogeographical Society 9: 41-45.

Reynolds, J.D. (1990) Ecological relationships of peatland invertebrates. In: Doyle, G.J. (ed.), Ecology and Conservation of Irish Peatlands. Royal Irish Academy, Dublin. pp. 135-143.

Rydin, H. & Jeglum, J.K. (2006) The biology of peatlands. Oxford University Press, New York.

Schouten, M.G.C. (1984) Some aspects of the ecogeographical gradient in the Irish ombrotrophic bogs, paper presented to 7th Int. Peat Congress, Dublin, vol. 1, pp. 414-432, The International Peat Society, Helsinki.

Schouten, M.G.C. (ed.) (2002) Conservation and Restoration of Raised Bogs – geological, hydrological and ecological studies. Dúchas – The Heritage Service of the Department of the

Environment and Local Government, Ireland; Staatsbosheheer, The Netherlands; and The Geological Survey of Ireland. pp. 220.

Smart, S.M., Robertson, J., Shield, E.J. & van de Poll, M.H. (2003) Locating eutrophication effects across British vegetation between 1990 and 1998. Global Change Biology 9: 1763-1774.

Turunen, J., Roulet, N.T., Moore, T.R. & Richard, P.J.H. (2004) Nitrogen deposition and increased carbon accumulation in ombrotrophic peatlands in eastern Canada. Global Biogeochemical Cycles. 18 (3): GB3002.

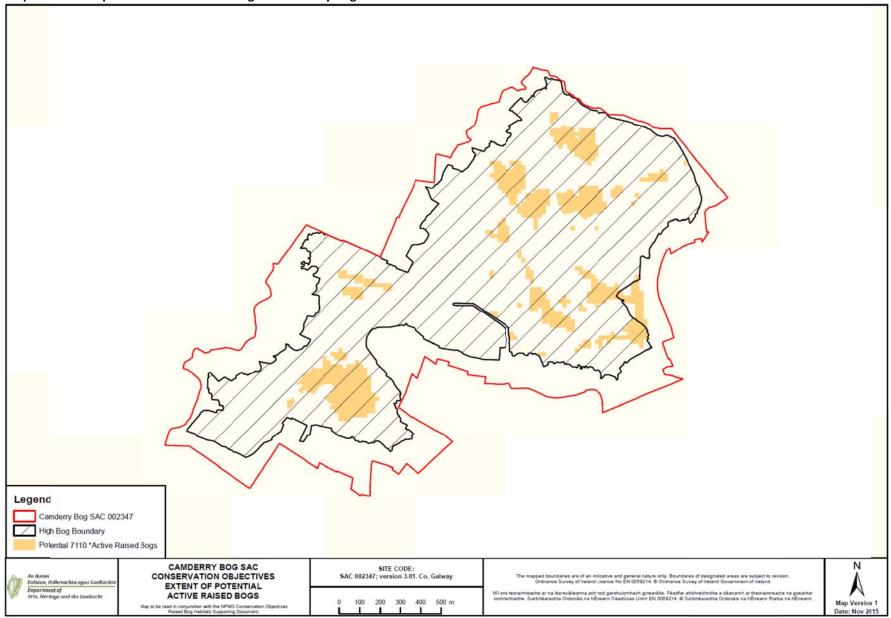
Van Duinen G.A. (2013) Rehabilitation of aquatic invertebrate communities in raised bog landscapes. PhD thesis, Radboud University Nijmegen, the Netherlands.

Verberk, W.C.E.P. (2008) Matching species to a changing landscape – Aquatic macroinvertebrates in a heterogeneous landscape. PhD thesis, Radboud University, Nijmegen.

Wilson, H.J. (1990) Birds of raised bogs. pp. 29-36. In: Cross, J. (ed.) The Raised Bogs of Ireland, their ecology, status and conservation. Report to the Minister of State at the Department of Finance. The Stationery Office, Dublin.

Wisdom, R. & Bolger, T. (2011) Chapter 2.4 - Terrestrial invertebrate biodiversity. pp. 103-121 In: Renou-Wilson, F. (ed.) BOGLAND: Sustainable Management of Peatlands in Ireland. Environmental Protection Agency, Wexford.

Map 1: Extent of potential active raised bog on Camderry Bog.



Map 2: Distribution of raised bog ecotopes on Camderry Bog. Legend Camderry Bog SAC 002347 / High Bog Boundary **Active Raised Bogs Ecotopes** Central ecotope Soaks / active flush Sub-central ecotope SITE CODE: SAC 002347; version 3.01. Co. Galway CAMDERRY BOG SAC An Roinn
Eulaien, Oidhreachta agus Gaeltachta
Department of
Arts, Heritage and the Gaeltacht The mapped boundaries are of an indicative and general nature only. Boundaries of designated areas are subject to revision.

Ordnance Survey of Ireland Licence No EN 0059214. © Ordnance Survey of Ireland Government of Ireland. CONSERVATION OBJECTIVES ACTIVE RAISED BOGS Nil sia teorainneacha ar na léarscáileanna ach nod garshuíomhach ginearáita. Féadfarathbhreithnithe a déanamh ar theorainneacha na gceantar comnarthaibhe. Suirbhéarachta Ordonáis na hÉireann Ceadúnas Umh EN 0050214. © Suirbhéarachta Ordonáis na hÉireann Riatas na hÉireann ECOTOPES 0 100 200 300 400 500 m Map Version 1 Map to be read in conjunction with the NPWS Conservation Objectives D

Map 3: Digital elevation model and drainage patterns at Camderry Bog.

